

Removal of selected pharmaceutical compounds from water by an organic polymer resin

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Received 14 July 2008; revised 05 February 2009; accepted 06 February 2009

This study presents adsorption performances for Carbamazepine, Propyphenazone, and Sulfamethoxazole using polymer resin, Lewatit VP OC 1163. Adsorption capacities followed: Carbamazepine > Sulfamethoxazole > Propyphenazone. Lewatit VP OC 1163 showed a larger adsorption for pharmaceuticals having low solubility, with Freundlich adsorption parameter, K_p , reaching 46.068 [(mgg^{-1}) (mg^{-1})] for Carbamazepine (solubility, 179 mg^{-1}). SAC_{254} (Spectral Absorption Coefficient at 254 nm) can be used as a control parameter for pharmaceutical adsorption studies.

Keywords: SAC_{254} , Adsorption, Freundlich isotherm, Pharmaceuticals, Polymer resin

Introduction

Pharmaceutical compounds (approx. 3000), approved as constituents for medicinal products in European Union¹, are inherently biologically active and can have unforeseen adverse effects on nontarget ecological species when released into environment^{2,3}. Wastewater treatment studies⁴⁻⁸ have shown that pharmaceuticals are released directly into environment. Three pharmaceuticals [Carbamazepine (CBZ), Propyphenazone (PP) and Sulfamethoxazole (SMX)] are used worldwide, with a production volume estimated to be many hundreds of tons annually, and are present in environment all over the world⁹⁻¹⁷. An effective treatment technology for removal of pharmaceuticals is by adsorption onto activated carbons¹⁸⁻²⁰. Polymer resins are becoming more common in wastewater treatment due to low costs, easy regeneration, and selective removal of pollutants²¹⁻²³. Lewatit VP OC 1163, a microporous, hypercrosslinked adsorber resin with good chemical and mechanical stability, has been characterized as having very high adsorption capacity²⁴.

This study assesses potential use of Lewatit VP OC 1163 as an adsorbent, both in batch and in column experiments, for removal of pharmaceuticals (CBZ, PP, and SMX) from aqueous solution.

Materials and Methods

Materials

Carbamazepine [5H-Dibenz(b,f)azepine-5-carboxamide), Propyphenazone (4-isopropyl-1,5-dimethyl-2phenyl-1,2-dihydro-3H-pyrazol-3-one], and Sulfamethoxazole [4-amino-N-(5-methylisoxazol-3-yl)-benzene sulfonamide] were obtained^{25,26} in e"99% purity from pharmaceutical-manufacturing units in Istanbul, Turkey. Stock solutions were prepared by dissolving pharmaceuticals in distilled water. Experiments were carried out at pH values of pharmaceuticals in solutions as follows: CBZ, 5.75-6.07; PP, 6.17-6.79; and SMX, 5.12-5.51.

A commercial organic polymer resin (Lewatit VP OC 1163) obtained from Bayer AG was used as an adsorbent. Lewatit VP OC 1163, a microporous, brown-red, translucent, hypercrosslinked adsorber resin without functional groups, has following properties²⁷: mean bead size, 0.45-0.55 mm; density, 1.09 g^{-1} ; surface area, >1300 m^2g^{-1} ; pore volume, 0.6-0.8 cm^3g^{-1} ; pore diam, 0.5-10 nm, and Zeta Potential (measured by Brookhaven Instruments BIC 90 Plus at pH 2-12) 0 mV.

Adsorption Studies

Batch Experiments

Batch experiments were carried out at $25 \pm 2^\circ\text{C}$ in a thermostated orbital shaker at an agitation speed of 150 rpm. Pharmaceutical solution (100 ml) was added to a

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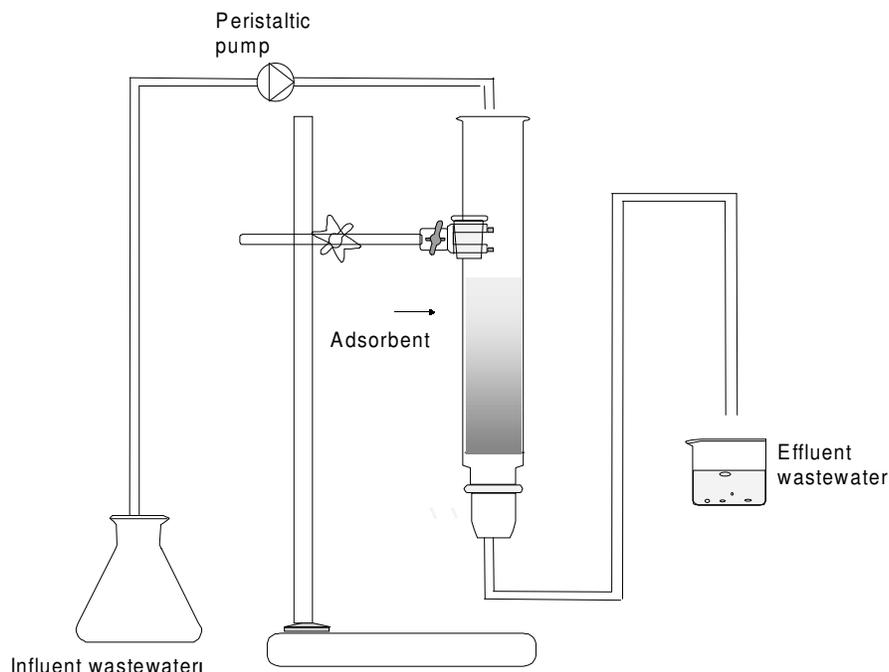


Fig. 1—Schematic representation of fixed bed column experimental setup

standard amount of adsorbent in stoppered Erlenmeyer flask. Blanks using distilled water were prepared to assess loss of solutes to flask components during adsorption tests. Results from duplicate flasks showed that average solution phase concentrations were within 96-102% of respective initial concentrations of same solutions analyzed in a similar manner. Hence, loss of compound was negligible. Supernatants were filtered through 0.45 μ m Whatman filter paper to determine contaminant concentration by two methods, KMnO_4 Demand²⁸ and SAC_{254} ²⁹ (Spectral Absorption Coefficient at 254 nm). Adsorption studies from aqueous solutions involve measurement of adsorbate concentration as a function of time, and amount of adsorbent used. SAC_{254} method was used for pharmaceutical adsorption studies. Spectrophotometric measurements were carried out using a Jenway 6105 UV-Vis spectrophotometer. Pharmaceutical amounts were determined from initial and final concentrations of solution. KMnO_4 Demand and SAC_{254} data were converted into concentration data using calibration relations for each pharmaceutical. Adsorbed amounts of pharmaceuticals were calculated as³⁰

$$q_e = (C_0 - C_e) \frac{V}{W} \quad \dots(1)$$

where q_e is amount of pharmaceutical adsorbed on adsorbent, C_0 is initial pharmaceutical concentration, C_e is equilibrium concentration of pharmaceutical solution, V is volume of pharmaceutical solution used and W is weight of adsorbent used. Removal efficiency, R , has also been calculated from batch experiments as

$$R = [(C_0 - C_e)/C_0].100 \quad \dots(2)$$

Column Experiments

Column tests were carried out using a column of glass (diam, 2 cm; ht, 45 cm). Column (Fig. 1) was packed with a 10 ml of adsorbents (12.7-13.2 cm ht). Glass wool was used as a support for adsorbents. Adsorbate solution was fed by a peristaltic pump (ISMATEC, MV-CA4). Column was charged in downflow mode with a volumetric flow rate [1.5 ml min^{-1} ($\sim 5 \text{ m}^3/\text{m}^2/\text{h}$)]. Pharmaceutical solutions with initial concentrations of 20 mg l^{-1} and 50 mg l^{-1} were used. Samples were collected at different time intervals and contaminant concentrations were determined by SAC_{254} . Temperature (T) was maintained at $25 \pm 2^\circ\text{C}$, and all studies were conducted at existing pH of solutions, which in turn depend on their concentration. To facilitate calculation of fixed bed adsorption capacity, breakthrough bed volume is often fixed at 50% or 10% of inlet concentration according to target quality of final effluent. Loading of adsorbents was continued until breakthrough

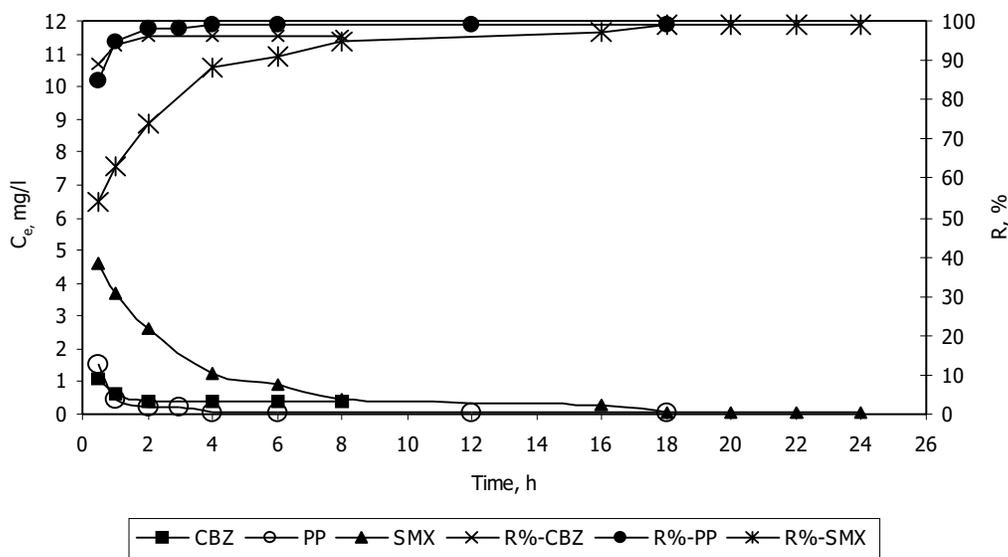


Fig. 2—Effect of contact time on pharmaceutical adsorption by VP OC 1163 ($C_0 = 10 \text{ mg l}^{-1}$, $T = 25^\circ\text{C}$, 150 rpm , $W = 0.1 \text{ g}$, $V = 20 \text{ ml}$)

was 10% of feed concentration. Quantity of adsorbed contaminant at breakpoint (q_b in mg g^{-1}) can be obtained as³¹

$$q_b = \frac{Q_v t_{10\%} C_0}{m_c} \dots(3)$$

where $t_{10\%}$ is service time (min) at outlet concentration 10% of inlet concentration, C_0 is inlet pharmaceutical concentration (mg l^{-1}), Q_v is effluent flow rate (l min^{-1}), and m_c is adsorbent amount (g).

Results and Discussion

Batch Experiments

Experiments were conducted for various time intervals to determine duration required to reach adsorption equilibrium (Fig. 2). Adsorption increased with increasing contact time for all pharmaceutical solutions, due to a large number of vacant surface sites are available for adsorption during initial stage, and after a lapse of time, remaining vacant surface sites are difficult to be occupied because of the repulsive forces between solute molecules in solid and bulk phases³². Removals for CBZ and PP were 89% and 85% within 1½ h and then gradually increased up to 96% and 99% in 2 h and 4 h, respectively. Because of such a quick sorption rate, it can be inferred that in this instance, process is predominantly controlled by film diffusion, and this is triggered by high concentration difference between bulk solution and solid phase³³.

Compared with other pharmaceutical solutions, a slow (gradual) uptake of SMX and establishment of

equilibrium after a long period (18 h) were observed. Most uptake occurred within first 8 h. Adsorption appeared to be governed by two transport processes. During first stage, SMX was rapidly adsorbed on organic polymer resin particles for 6-8 h. In second stage, slower migration of pharmaceutical to less accessible sites within resin matrix could be observed for a period totaling 18 h. Therefore, a large amount of SMX was expected to be progressively adsorbed by adsorbent matrices as contact time increases³⁴. Slow uptake of adsorbates and establishment of equilibrium over a long period indicate strong chemical binding of adsorbate with adsorbent³².

An increase in adsorbent dosage increased percentage removal of pharmaceuticals. Increase in R with increase of adsorbent dose can be attributed to increased number of sites available for adsorption³⁵. Lewatit VP OC 1163 had a greater adsorption affinity for CBZ (0.1 g/20 ml) and SMX (0.1 g/20 ml) than that for PP (0.4 g/20 ml) (Fig. 3). Initial concentration affects both adsorption capacity of adsorbent and adsorption rate. With increasing concentrations, R values from SAC_{254} and KMnO_4 Demand measurements increased for CBZ and PP and decreased for SMX (Table 1).

Freundlich model is based on a monolayer adsorption and is expressed as

$$q_e = K_f C_e^n \dots(4)$$

where K_f and n are noncompetitive Freundlich constant characteristics of the system. K_f and n are indicators of adsorption capacity and adsorption intensity,

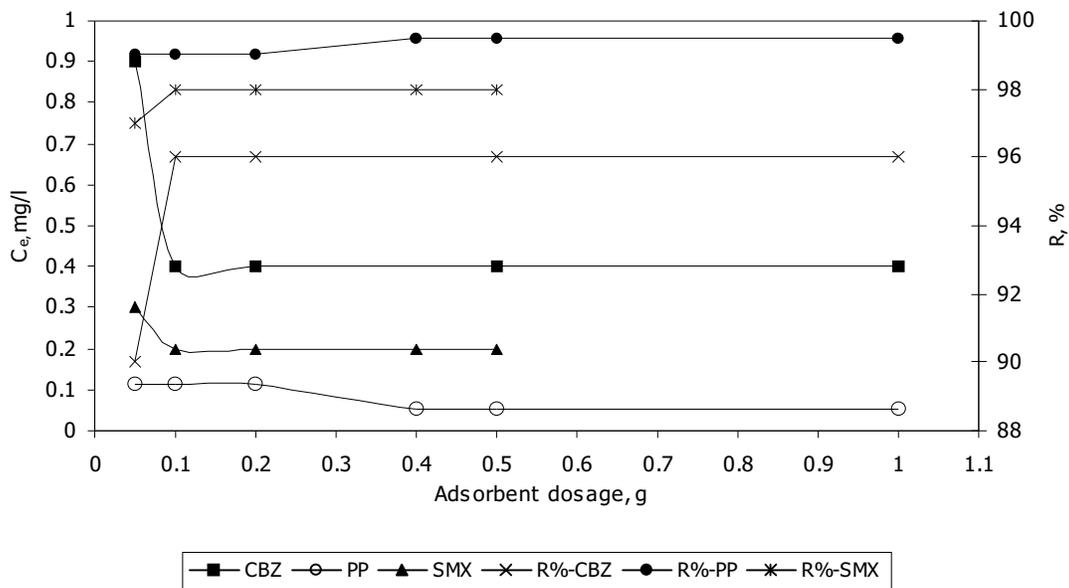


Fig. 3—Effect of adsorbent dosage on pharmaceutical adsorption by VP OC 1163 (C₀ = 10 mg l⁻¹, T = 25°C, 150 rpm, V = 20 ml)

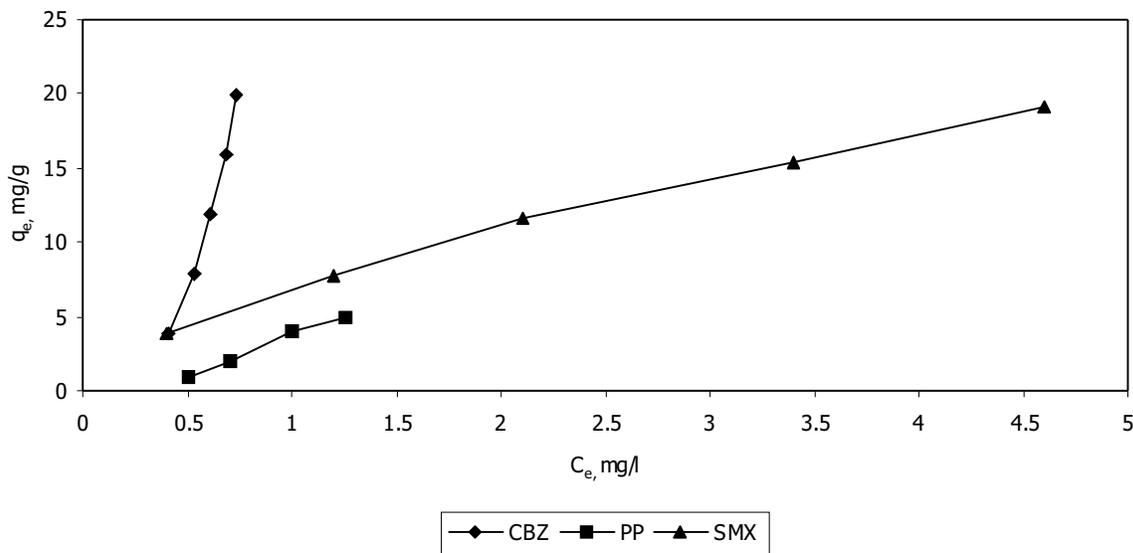


Fig. 4—Adsorption isotherms of pharmaceuticals onto VP OC 1163 at 25°C (obtained from SAC₂₅₄ measurements)

respectively. Langmuir model, originally developed to represent monolayer sorption on a set of distinct localized sorption sites, is expressed as

$$q_e = \frac{Q^0 b C_e}{1 + b C_e} \quad \text{(Non-linear form)} \quad \dots(5)$$

$$\frac{C_e}{q_e} = \frac{1}{Q^0 b} + \frac{C_e}{Q^0} \quad \text{(Linear form)} \quad \dots(6)$$

where Q⁰ is maximum amount of adsorbate per unit weight of adsorbent that forms a complete monolayer on the surface and b is a constant related to affinity of binding sites³⁰.

Experimental equilibrium data of pharmaceuticals adsorbed on Lewatit VP OC 1163 were obtained using SAC₂₅₄ (Fig. 4) and KMnO₄ demand measurements (Fig. 5). Liquid adsorption isotherms have been classified into four categories (S, L, H, and C types) with subdivisions for each type³⁶⁻³⁷. Isotherms obtained from both parameters [SAC₂₅₄ and KMnO₄ demand]

Table 1—Effect of initial pharmaceutical concentrations on the amount of adsorbed pharmaceuticals and corresponding removal efficiencies

C ₀ mg l ⁻¹	R%					
	Carbamazepine		Propyphenazone		Sulfamethoxazole	
	SAC ₂₅₄	KMnO ₄ demand	SAC ₂₅₄	KMnO ₄ demand	SAC ₂₅₄	KMnO ₄ demand
20	98	99	97.5	n d	98	95.8
40	98.7	99.3	98.3	n d	97	95.3
60	99	99.4	-	n d	96.5	95.1
80	99.1	99.4	98.8	n d	95.8	95.3
100	99.3	99.5	98.8	n d	95.4	95

* n d: not determined

Table 2—Estimated parameters for Freundlich isotherm model at 25°C

Pharmaceuticals	SAC ₂₅₄				KMnO ₄ demand			
	K _f	n	R ²	Σ %	K _f	n	R ²	Σ %
Carbamazepine	46.068	2.7862	0.9984	1.18	11.792	1.3668	0.9922	0.91
Propyphenazone	3.573	1.7979	0.9856	1.73	n d	n d	n d	n d
Sulfamethoxazole	7.04	0.6474	0.9993	0.2	1.103	0.9054	0.9976	0.45

* n d: not determined

showed same shape. In studied concentration ranges, isotherms of CBZ and PP may be classified as belonging to S type, suggesting multilayer adsorption and vertical orientation of adsorbed pharmaceutical molecules on adsorbent surface, with availability of new sites to solvent as adsorption proceeds. It also indicates that solvent and solute compete with each other for adsorption sites on adsorbent surface. Upward nature of curves shows that, after complete adsorption on adsorbent surface, adsorbate molecules attract each other and become associated to a considerable extent³⁸. According to the slope of initial portion of curves, isotherms of SMX may be classified as L type, indicating high affinity of adsorbent for adsorbate. This shows that no strong competition occurs for adsorption sites between solvent molecules (distilled water) and adsorbate molecules (pharmaceutical molecules). However, as active sites of adsorbent become saturated, adsorption of new molecules occurs with greater difficulty. Adsorption reaches a constant saturation value³⁹.

Estimated parameters of Freundlich model (Table 2) are given with average percentage errors. Langmuir equation did not reproduce equilibrium data satisfactorily for CBZ and PP, and estimated Langmuir parameters were negative. Langmuir isotherm values of

KMnO₄ Demand measurements of PP batch experiments were not calculated due to lack of data because PP removal was below detection limit. Average percentage errors were calculated as⁴⁰

$$\epsilon \% = \frac{\sum_{i=1}^N \frac{|q_{e,i,exp} - q_{e,i,calc}|}{q_{e,i,exp}}}{N} \cdot 100 \quad \dots(7)$$

where N is number of measurements, ‘exp’ and ‘calc’ represent experimental and calculated q_e values.

Fit of experimental data obtained from both SAC₂₅₄ and KMnO₄ Demand measurements for Freundlich equation is good for CBZ and PP in studied concentration range. Correlation coefficients showed that fit to Freundlich model is excellent. Average percentage error values calculated for fit of isotherm data of three pharmaceuticals to Freundlich equation are in accordance with correlation coefficients. For SMX adsorption, both Langmuir (R²= 0.9885) and Freundlich model (R² = 0.9993) generated a satisfactory fit of experimental data. Applicability of both Langmuir and Freundlich isotherms to the system may indicate that both monolayer and heterogeneous surface conditions

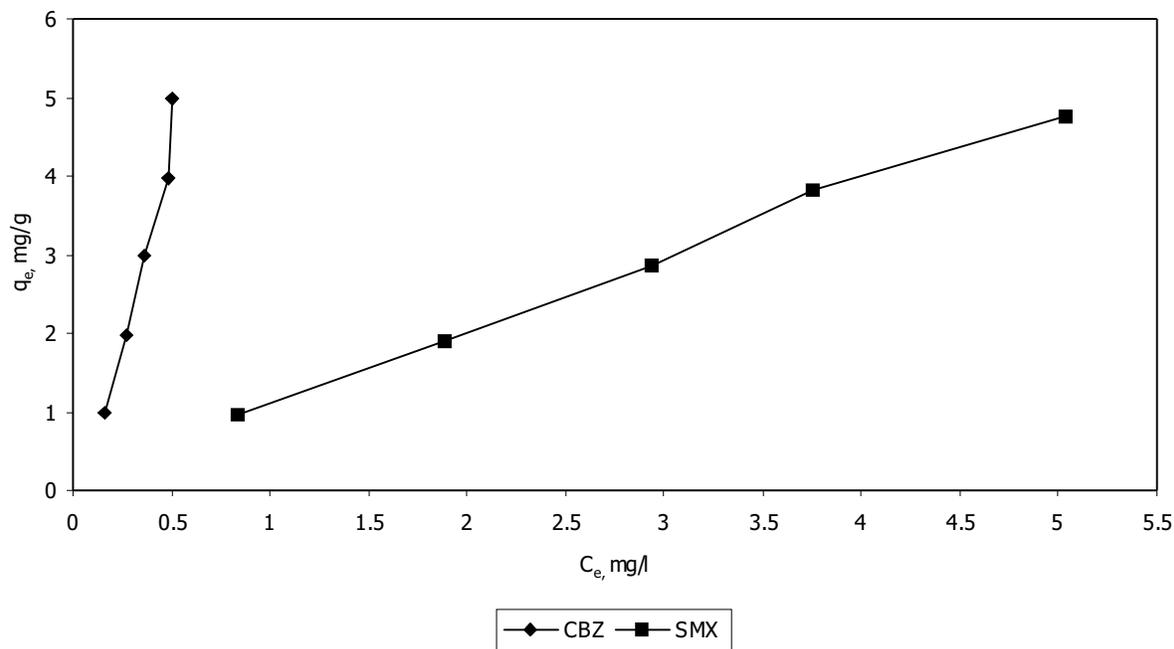


Fig. 5—Adsorption isotherms of pharmaceuticals onto VP OC 1163 at 25°C (obtained from KMnO_4 demand measurements)

Table 3—Adsorption capacity of Lewatit VP OC 1163 at different initial pharmaceutical concentrations (BV = 10 ml, Z = 12.7-13.2 cm, $m_c = 6.7391$ g, $v = 5$ mh^{-1} , and $Q_v = 6.6$ mlmin^{-1})

Pharmaceuticals	Initial concentration, C_0 mg l^{-1}	Treated volume, v_b ml^a	Breakthrough point, t_b min^a	Adsorption bed capacity, q_b mg g^{-1b}
Carbamazepine	20	68700	10410	203.9
	50	30200	4576	224.08
Propyphenazone	20	57100	8652	169.47
	50	28000	4242	207.72
Sulfamethoxazole	20	64200	9727	190.53
	50	30100	4561	223.34

^aObtained experimentally

^bCalculated according to Eq. (3)

exist under experimental conditions used⁴². However, higher values of correlation coefficient and lower percentage errors for Freundlich model, compared with Langmuir ($\% = 0.2$ and 47.5 , respectively) indicated that Freundlich model gave a better fit of the data.

Adsorbed amounts and K_f values of three pharmaceutical solutions (PP < SMX < CBZ) indicate higher adsorption capacity for CBZ. In general, adsorption increases with decreasing water solubility and increasing compound hydrophobicity, as represented by octanol-water partitioning coefficient ($\log K_{ow}$). For CBZ, PP and SMX, main cause for variation in K_f was

varying adsorbate solubility. This result is in agreement with Lundelius rule. A lower adsorption capacity was obtained for highly soluble compounds (PP and SMX) in comparison with less soluble compound (CBZ).

At a pH value between pK_a (dissociation constant) values of compounds, pharmaceuticals exist predominantly as neutral species. Nature of solid surface, either hydrophobic or hydrophilic, and electrical interactions play an important role in adsorption kinetics of contaminant at solid-liquid interface. In deionized water (pH 2, 4, 6, 8, 10, and 12), ζ -potentials of organic polymer resin particles were all zero. Therefore, resin

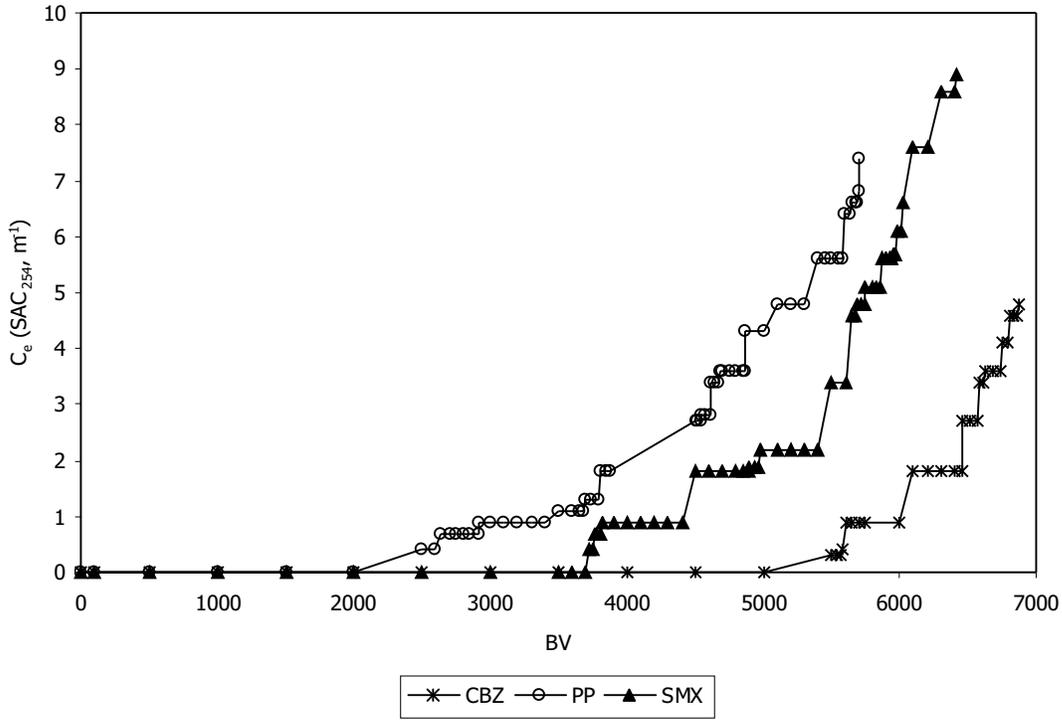


Fig. 6—Breakthrough curves of 20 mg l^{-1} pharmaceutical solutions using Lewatit VP OC 1163 ($V = 10$ ml, $Z = 12.7$ – 13.2 cm)

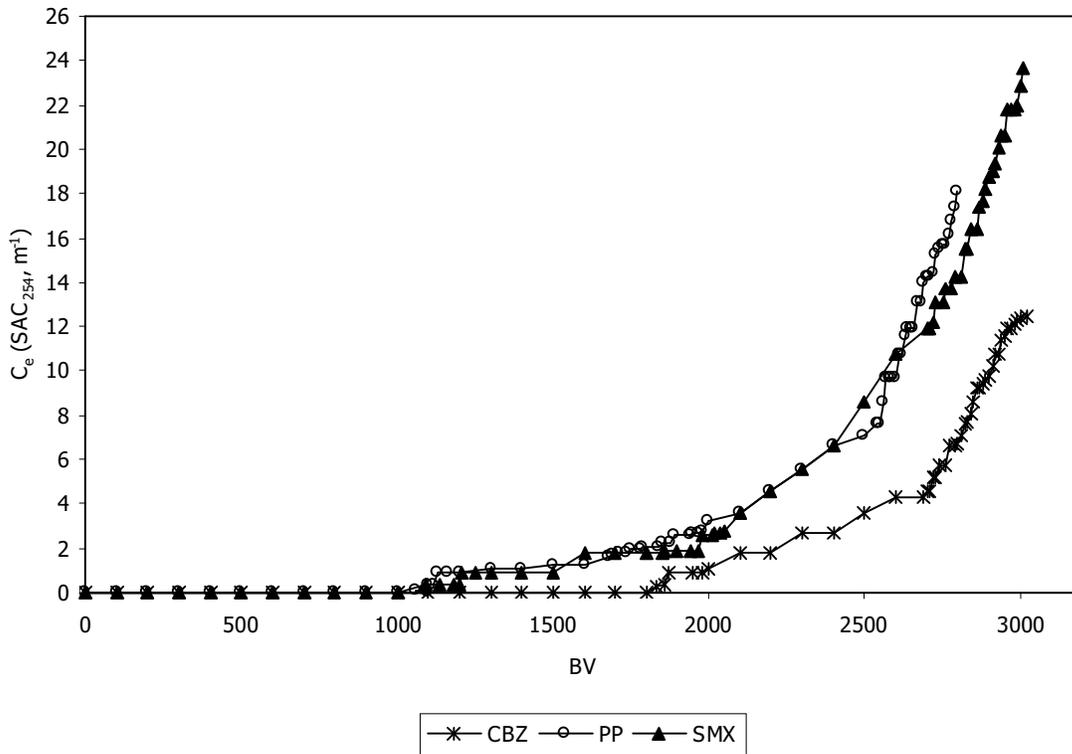


Fig. 7—Breakthrough curves of 50 mg l^{-1} pharmaceutical solutions using Lewatit VP OC 1163 ($V = 10$ ml, $Z = 12.7$ – 13.2 cm)

surface was uncharged in deionized water within a broad pH range. In this context, it could be assumed that electrostatic interaction might not significantly contribute to adsorption of pharmaceuticals. Lower the hydrophobicity ($\log K_{ow}$) of a compound, higher is the expected adsorption onto organic polymer resin. Because organic polymer is a hydrophilic (weak polar), higher adsorption capacity was expected in the case of relatively hydrophilic (polar) PP. But estimated values of Freundlich parameter showed that adsorptive capacity of CBZ was higher than that of PP and SMX ($K_f = 46.07, 3.573, \text{ and } 7.04$, respectively). However, adsorption of SMX was higher than that of PP despite $\log K_{ow}$ of SMX is lower than that of PP. Thus, for the studied system, no simple relationship was found between hydrophobicity of pharmaceuticals and adsorption affinity toward hydrophilic surface of the adsorbent. Freundlich fitting with $n > 1$ for CBZ and PP would indicated conformity of the data to a multilayer formation on adsorbent surface. For SMX ($n < 1$) implies that organic polymer resin-SMX systems correspond to a favorable adsorption process.

Column Experiments

Only SAC_{254} measurements could be used for evaluation of column experiments because of low outlet column concentrations. Table 3 shows treated volume (V_b), breakthrough time (t_b), and quantity of adsorbed contaminant at breakthrough point (q_b). Good adsorbability of CBZ on Lewatit VP OC 1163 could be confirmed by column experiments. 10% of initial CBZ concentration was reached after 68700 ml (6870 BV (bed volume) for 20 mg l^{-1} (Fig. 6) and 30200 ml (3020 BV) for 50 mg l^{-1} (Fig. 7). PP and SMX were generally less adsorbed than CBZ, can be attributed to physicochemical properties of solutes, mainly solubility. When initial concentrations were increased, breakthrough point was reached more quickly. A rise in initial pharmaceutical concentration reduced volume that was treated before 10 % of initial concentration was reached. S-shaped breakthrough curves obtained from column studies showed that less than half of adsorbent capacity was used until breakthrough point.

Conclusions

Experimental data of adsorption capacities under optimized conditions were best fitted to Freundlich model. Estimated values of Freundlich parameter showed that adsorptive capacity of CBZ was higher than

that of PP and SMX ($K_f = 46.068, 3.573, \text{ and } 7.04$, respectively), implying that adsorbate solubility has most impact on adsorption. After 2-4 h contact, a steady-state approximation was assumed and quasi-equilibrium was accepted for CBZ and PP. This is a very important aspect for pharmaceutical industry because less contact time could have a favorable impact on process economics. Good adsorbability of CBZ on Lewatit VP OC 1163 could be confirmed by column experiments. Relatively high solubilities of PP and SMX are responsible for short breakthrough times observed using Lewatit VP OC 1163 column. Breakthrough curves (S-shaped) show that less than half the adsorbent capacity was used until breakthrough point. SAC_{254} is a useful parameter for determining pharmaceutical concentration in aqueous solutions with advantage that no chemicals are used and measuring time is shorter compared to other organic matter parameters. Organic polymer resin is found effective for removing pharmaceuticals.

References

- 1 Laville N, Ait-Aissa S, Gomez E, Casellas C & Porcher J M, Effects of human pharmaceuticals on cytotoxicity, EROD activity and ROS production in fish hepatocytes, *Toxicology*, **196** (2004) 41-55.
- 2 Breton R & Boxall A, Pharmaceuticals and personal care products in the environment: regulatory drivers and research needs, *QSAR Comb Sci*, **22** (2003) 399-409.
- 3 Sanderson H, Johnson D J, Reitsma T, Brain R A, Wilson C J & Solomon K R, Ranking and prioritization of environmental risks of pharmaceuticals in surface waters, *Regul Toxicol Pharm*, **39** (2004) 158-183.
- 4 Ternes T A, Occurrence of drugs in german sewage treatment plants and rivers, *Water Res*, **32** (1998) 3245-3260.
- 5 Daughton C G & Ternes T A, Pharmaceuticals and personal care products in the environment: Agents of subtle change?, *Environ Health Persp Suppl*, **107** (1999) 907-938.
- 6 Carucci A, Cappai G & Piredda M, Biodegradability and Toxicity of Pharmaceuticals in Biological Wastewater Treatment Plants, *J Environ Sci Heal A*, **41** (2006) 1831-1842.
- 7 Heberer T, Occurrence, fate and removal of pharmaceutical residues in the aquatic environment : a review of recent research data, *Toxicol Lett*, **131** (2002) 5-17.
- 8 Skavanis C, Groundwater disaster in Puerto Rico—the need for environmental education, *J Environ Hlth*, **62** (1999) 29-35.
- 9 Halling-Sorensen B, Nors N S, Lanzky P F, Ingerslev F, Holten Lützhof H C & Jorgensen S E, Occurrence, fate and effects of pharmaceutical substances in the environment—A review, *Chemosphere*, **36** (1998) 357-393.
- 10 Heberer T, Schmidt-Bäumler K & Stan H J, Occurrence and distribution of organic contaminants in the aquatic system in Berlin. Part I: Drug residues and other polar contaminants in Berlin surface and groundwater, *Acta Hydroc Hydrob*, **26** (1998) 272-278.

- 11 Wollenberger L, Halling-Sorensen B & Kusk K O, Acute and chronic toxicity of veterinary antibiotics to *Daphnia magna*, *Chemosphere*, **40** (2000) 723-730.
- 12 Hirsch R, Ternes T, Haberer K & Kratz K, Occurrence of antibiotics in the aquatic environment, *Sci Total Environ*, **225** (1999) 109-118.
- 13 Carballa M, Omil F, Lema J M, Llopart M, García-Jares C, Rodríguez I, Gómez M & Ternes T, Behavior of pharmaceuticals, cosmetics and hormones in a sewage treatment plant, *Water Res*, **38** (2004) 2918-2926.
- 14 Xu W, Zhang G, Li X, Zou S, Li P, Hu Z & Li J, Occurrence and elimination of antibiotics at four sewage treatment plants in the Pearl River Delta (PRD), South China, *Water Res*, **41** (2007) 4526-4534.
- 15 Göbel A, McArdell C S, Joss A, Siegrist H & Giger W, Removal of pharmaceuticals and fragrances in biological wastewater treatment, *Water Res*, **39** (2005) 3139-3152.
- 16 Strenn B, Clara M, Gans O & Kreuzinger N, Carbamazepine, diclofenac, ibuprofen and bezafibrate—investigations on the behaviour of selected pharmaceuticals during wastewater treatment, *Water Sci Technol*, **50** (2004) 269-276.
- 17 Zuehlke S, Duennbier U & Heberer T, Investigation of the behavior and metabolism of pharmaceutical residues during purification of contaminated ground water used for drinking water supply, *Chemosphere*, **69** (2007) 1673-1680.
- 18 Fuerhacker M, Dürauer A & Jungbauer A, Adsorption isotherms of 17 ²-estradiol on granular activated carbon (GAC), *Chemosphere*, **44** (2001) 1573-1579.
- 19 Melillo M, Phillips G J, Davies J G, Llyod A W, Tennison S R, Kozynchenko O P & Mikhalovsky S V, The effect of protein binding on ibuprofen adsorption to activated carbons, *Carbon*, **42** (2004) 565-571.
- 20 Snyder S A, Adham S, Redding A M, Cannon F S, DeCarolis J, Oppenheimer J, Wert E C & Yoon Y, Role of membranes and activated carbon in the removal of endocrine disruptors and pharmaceuticals, *Desalination*, **202** (2007) 156-181.
- 21 Chen P H, Adsorption of organic compounds in water using a synthetic adsorbent, *Environ Int.*, **23** (1997) 63-73.
- 22 Kaya Y, Vergili I, Gönder Z B & Barlas H, Investigation of organic matter removal from waters with adsorption polymers, *Fresen Environ Bull*, **15** (2006) 437-440.
- 23 Dutta M, Dutta N N & Bhattacharya K G, Aqueous phase adsorption of certain beta-lactam antibiotics onto polymeric resins and activated carbon, *Sep Purif Technol*, **16** (1999) 213-224.
- 24 BAYER AG, Lewatit VP OC 1163 Technical Brochure (2002).
- 25 Nghiem L D & Hawkes S, Effects of membrane fouling on the nanofiltration of pharmaceutically active compounds (PhACs): Mechanisms and role of membrane pore size, *Sep Purif Technol*, **57** (2007) 176-184.
- 26 Mersmann P, *Transport und Sorptionsverhalten der Arzneimittelwirkstoffe Carbamazepin, Clofibrinsäure, Diclofenac, Ibuprofen und Propyphenazon in der wassergesättigten und -ungesättigten Zone*, Ph D Thesis, Institut für Angewandte Geowissenschaften der Technischen Universität Berlin, Germany, 2003.
- 27 Alkan M, Karadas M, Dogan M & Demirbas Ö, Adsorption of CTAB onto perlite samples from aqueous solutions, *J Colloid Interf Sci*, **291** (2005) 309-318.
- 28 DIN EN ISO 8467, *Wasserbeschaffenheit, Bestimmung des Permanganat-Index*, Beuth Verlag, Berlin, Germany, 1995.
- 29 DIN 38404-C3, *Deutsche Einheitsverfahren zur Wasser-, Abwasser- und Schlammuntersuchung- physikalisch-chemische Kenngrößen (Gruppe C)-Teil 3: Bestimmung der Absorption im Bereich der UV-Strahlung; Spektraler Absorptionskoeffizient (C3)*, Beuth Verlag, Berlin, Germany, 2003.
- 30 Sontheimer H, Frick BR, Fettig J, Hörner G, Hubele C & Zimmer G, *Adsorptionsverfahren zur Wasserreinigung* (DVGW-Forschungsstelle am Engler-Bunte-Institut der Universität Karlsruhe (TH), Karlsruhe) 1985, 120-139.
- 31 Taty-Costodes V C, Fauduet H, Porte C & Ho YS, Removal of lead (II) ions from synthetic real effluents using immobilized *Pinus sylvestris* sawdust: Adsorption on a fixed-bed column, *J Hazard Mater*, **B123** (2005) 135-144.
- 32 Malli I D, Srivastava V C & Agarwalli N K, Removal of Orange-G and Methyl Violet dyes by adsorption onto bagasse fly ash—kinetic study and equilibrium isotherm analyses, *Dyes Pigments*, **69** (2006) 210-223.
- 33 Adak A, Bandyopadhyay M & Pal A, Removal of anionic surfactant from wastewater by alumina: a case study, *Colloid Surface A*, **254** (2005) 165-171.
- 34 Zeledon-Toruno ZC, Lao-Luque C, De Las Heras F X C & Sole-Sardans M, Removal of PAHs from water using an immature coal (leonardite), *Chemosphere*, **67** (2007) 505-512.
- 35 Gupta S, Pal A, Ghosh P K & Bandyopadhyay M, Performance of waste activated carbon as a low-cost adsorbent for the removal of anionic surfactant from aquatic environment, *J Environ Sci Hlth A*, **38** (2003) 381-397.
- 36 Giles C H, Macewan T H, Nakhwa S N & Smith D, Studies in adsorption. Part XI. A system of classification of solution adsorption isotherms, and its use in diagnosis of adsorption mechanisms and in measurement of specific surface areas of solids, *J Chem Soc*, **111** (1960) 3973-3993.
- 37 Giles C H & Nakhwa S N, Studies in adsorption. Part XVI, The measurement of specific surface areas of finely divided solids by solution adsorption, *J Appl Chem*, **12** (1962) 266-273.
- 38 Singh D, Co-solvent effects on the adsorption of carbofuran by two indian soils, *Pesticide Mgmt Sci*, **56** (2000) 195-201.
- 39 Rodriguez-Cruz M S, Sanchez-Martin M J & Sanchez-Camazano M, A comparative study of adsorption of an anionic and a non-ionic surfactant by soils based on physicochemical and mineralogical properties of soils, *Chemosphere*, **61** (2005) 56-64.
- 40 Aksu Z & Kabasakal E, Batch adsorption of 2,4-dichlorophenoxy acetic acid (2,4-D) from aqueous solution by granular activated carbon, *Sep Purif Technol*, **35** (2004) 223-240.
- 41 Sotelo J L, Ovejero G, Delgado J A & Martinez I, Comparison of adsorption equilibrium and kinetics of four chlorinated organics from water onto GAC, *Water Res*, **36** (2002) 599-608.