

Soil Contamination in Nuclear Reactor Surrounding Areas in Savar, Bangladesh using Instrumental Neutron Activation Analysis Method

M. Shahabuddin¹, M. Dilder Hossain¹, S. M. Hossain², M. Monzurul Hoque², M. Mamun Mollah³, and M. A. Halim⁴

1- Department of Physics, Jahangirnagar University, Savar, Dhaka-1342, Bangladesh

2- Atomic Energy Research Establishment, Savar, P.O. Box. 3787, Dhaka-1000, Bangladesh

3-National Measurement Institute, Bertie Street, Port Melbourne, Victoria 3207, Australia

4-Department of Chemistry, Memorial University, St. John's, NL, A1B 3X7, Canada

mahalim@mun.ca

ABSTRACT

In this paper, the contamination level of fourteen elements (Al, Ca, Mg, V, Ti, Na, Mn, As, Br, La, Sm, Sc, U, Yb) from different soil samples of the nuclear reactor surrounding areas in Savar, Bangladesh was investigated with instrumental neutron activation analysis (INAA) method. Generally, neutron activation is most reliable tool for determining the concentration of the trace elements and heavy metals by measuring the gamma rays released while decaying of the radioisotopes created from the interaction of neutrons with atoms. In this study, the soil contamination level of all heavy metals and trace elements was compared to the range, mean, and median values of the World soil as well as with other studies. The concentration of K (1.65%), U (3.17 mg kg⁻¹), Sm (6.10 mgkg⁻¹) and Sc (10.53 mg kg⁻¹) is relatively higher than that of the World mean value. Furthermore, compared to world range value for soil revealed that the high concentration range is also observed for Al (3.87-9.39%), Na (0.37-0.76%) and Yb (0.11-5.36 mg kg⁻¹). However, contamination level of all elements was not evenly distributed in the studied areas. For instance, the pollution levels of U, Yb, Mn, and V were unevenly higher in the areas near to the overhead tank, decay tank, in the vicinity of the underground tank, cooling tower and reactor building area. Interestingly, the contamination level of potassium is high everywhere near the nuclear reactor installations.

Key words: Elemental analysis, neutron activation, heavy metals, nuclear reactor, trace element

1. Introduction

The soil, a main part of the terrestrial ecosystem, is a heterogeneous mixture of different organisms and minerals, organic, and organo-mineral substances present in three phases: solid, liquid and gas [1-2]. At the same time, soil is perhaps the most endangered component of our environment open to potential contamination by a variety of different pollutants arising from human activities such as nuclear, industrial, agricultural, etc., [3-5]. Although the trace elements in soil are very important for the quality of soil and environment, excessive level of trace elements can cause pollution of waters, toxicity in plants, foods and ultimately in animals and humans that feed upon them [6-8]. As a result, the limits of metal deposition rates to soils should take into account the general ecotoxicity, phytotoxicity, transfer to animals and risk to the human health. The trace metal concentrations in soils may be affected by the deposition of wastes released from various industries, nuclear research and

other related establishments. The erosive weathering, lead-base paints from the outer walls of buildings, land treatment of sewage sludge and natural sources of trace elements, e.g., windblown dust, forest wildfires, sea salt emission and vegetation may also contribute to the soil trace metal elevation [9]. Uncontrolled development of industry, agriculture and urbanization accelerates the input of heavy metals into the environment in many parts of the world [1].

Heavy metals have been the subject of particular attention because of their long-standing toxicity, mobility in the ecosystems and transfer into the food chains when specific thresholds have been exceeded [10-16]. Many procedures have been devoted to determining the sources, types, and degrees of heavy metal pollution in soil [17-20]. An important background for this kind of work requires the knowledge of geochemical baseline concentrations of elements. The trace element pollution depends on geochemical and biochemical properties of a given element. There is an increasing demand to measure background levels of inorganic materials in soils because of the need to determine whether a soil is contaminated or not.

In 1981, a research reactor was installed at Atomic Energy Research Establishment (AERE), Savar, Dhaka and used for training, research, and isotope production. Some other nuclear installations such as the decay tank, cooling tower, neutron generator and detection laboratory, isotope production laboratory, chimney, water treatment plant, and underground tank are also available nearby the research reactor. Although the reactor and the nearby installations are shielded well, there is still some probability of polluting the soil with heavy metals and other harmful substances. The geochemical baseline concentrations of metals in soil for the region are not available. The aim of the present work is to assess the contamination level of the soil samples taken at different points from the surrounding areas of the research reactor. The risk assessment of the present study is based on the comparison between the levels of contamination in the studied areas and the established guidelines or screening values [21, 22]. In addition, level of contamination in the soil samples was estimated by comparing with the world average levels and the other specified guidelines for the element in soils. It is more likely that the high content of some elements in the soils is a result of activities in the nuclear installations.

2. Materials and Methods

The soil samples were collected from the different locations of the surrounding areas of the TRIGA MARK II Research Reactor and other nearby nuclear installations of the Bangladesh Atomic Energy Commission (BAEC), Savar, Dhaka. Global Positioning System (GPS) technology was used to mark the sampling points (Table 1). The samples were collected in polyethylene bags, which were previously cleaned with dilute nitric acid, pure water, acetone, and then dried. The samples were at first dried naturally at room temperature in a clean room and then dried in an oven for 7 days at 50°C. The samples were then crushed into powers of 200 microns size particles in an agate mortar. About 50 mg of each sample was put in a clean polyethylene bag which was doubly heat-sealed, and later used for irradiation. For quality assurance purpose about 50 mg of three certified reference materials (CRM)

Table 1: Sampling points and location of investigated soil samples

Sample No.	Sampling points by GPS Longitude/ Latitude	Sample collection places surrounding the TRIGA MARK II Research Reactor	Sample No.	Sampling points by GPS Longitude/ Latitude	Sample collection places surrounding the TRIGA MARK II Research Reactor
RFS-1	N:23°57.122' E:90°16.787'	Close to decay tank	RFS-8	N:23°57.105' E:90°16.816'	Water treatment plant
RFS-2	N:23°57.108' E:90°16.805'	South side of Mehadi tree	RFS-9	N:23°57.082' E:90°16.784'	South west of reactor boundary
RFS-3	N:23°57.105' E:90°16.813'	East side of water tank plant	RFS-10	N:23°57.090' E:90°16.792'	West of Underground tank
RFS-4	N:23°57.091' E:90°16.799'	Near the over head tank	RFS-11	N:23°57.100' E:90°16.783'	Garden beside lemon tree
RFS-5	N:23°57.089' E:90°16.803'	Near the over head tank(South side of palm tree)	RFS-12	N:23°57.103' E:90°16.789'	West side cooling tower pipe
RFS-6	N:23°57.106' E:90°16.794'	North side of cooling tower	RFS-13	N:23°57.131' E:90°16.789'	North-East corner of reactor building
RFS-7	N:23°57.118' E:90°16.807'	Far from decay tank	RFS-14	N:23°57.112' E:90°16.781'	West side of reactor building

RFS: Reactor Facility Soil, GPS: Global Positioning System.

of International Atomic Energy Agency (IAEA) such as IAEA-soil-7, NBS-1633b Coal fly ash, and IAEA-SL-1 were also put in polyethylene bags for irradiation.

The samples and standards were simultaneously irradiated using 3MW neutron beam research reactor, TRIGA MARK II at the Bangladesh Institute of Nuclear Science and Technology (INST) under the Atomic Energy Research Establishment (AERE) of Bangladesh Atomic Energy Commission (BAEC). The short irradiation was performed using the pneumatic transfer system. Each sample and standard were irradiated for two minutes with 250 kw reactor power having a thermal neutron flux of 2.75×10^{13} neutrons per cm^2 per second. The long irradiation was performed in the rotary specimen rack (Lazy Susan). For long irradiation all the samples and standards were packed together in an irradiation vial and irradiated simultaneously for four hours with a reactor power of 250 kw and a thermal neutron flux of 2×10^{12} neutrons per cm^2 per second.

After irradiation the cover bag was removed and the bag the containing radioactive sample was again covered with a new clean polyethylene bag to avoid contamination during gamma-ray measurements. The activities of the irradiated samples and standards were measured nondestructively using a high purity germanium (HPGe) detector coupled with a pulse-height multi-channel (4,096 channels) analyzer. The γ -rays emitted from both the samples and standards were measured at the same geometry. The dead time of the detector was kept below 3%. The cooling time varied from 2 minutes to 72 h for the measurement of γ -rays depending on the irradiation (short and long) process. The counting time also varied with the irradiation process. Target and daughter nuclides, as well as other nuclear data pertinent to this study are presented in Table 2. In this study, the elemental concentrations

in the soil samples were calculated using the comparative method. The gamma peak analysis was performed using the software Hypermet PC version 5.12. The concentration was calculated manually in an Excel spread sheet as well as with the Peakgr-10.

Table 2: Decay data of the radionuclides and the contributing reactions

Element	Isotopic abundance I_γ (%)	Contributing reaction	Half-life	γ -ray energy (keV)
As	44.6%	$^{75}\text{As} (n, \gamma) ^{76}\text{As} \xrightarrow{\beta} ^{76}\text{Se}$	26.3 h	559.1
Br	83.2%, 72%	$[^{81}\text{Br} (n, \gamma) ^{82}\text{Br}]$	35.3h	776.5, 554.3
La	100%	$[^{139}\text{La} (n, \gamma) ^{140}\text{La}]$	40.3 h	818.5
Sm	28.3%	$[^{152}\text{Sm} (n, \gamma) ^{153}\text{Sm}]$	46.8 h	103.2
Sc	100%	$[^{45}\text{Sc} (n, \gamma) ^{46}\text{Sc}]$	3.41d	889, 1120
Yb	3.7%, 6%	$[^{176}\text{Yb} (n, 2n) ^{175}\text{Yb} + \gamma]$	4.2d	282.6, 396
U	12%, 18%		2.4 d	103.7, 277.6
Ti	95%	$[^{50}\text{Ti} (n, \gamma) ^{51}\text{Ti}]$	5.76 m.	320
Mg	100%	$[^{26}\text{Mg} (n, \gamma) ^{27}\text{Mg}]$	9.45 m	1014.4
V	100%	$[^{51}\text{V} (n, \gamma) ^{52}\text{V}]$	3.755 m	1434.2
Al	100%	$[^{27}\text{Al} (n, \gamma) ^{28}\text{Al}]$	2.246 m	1778.8
Ca	91.7%	$[^{48}\text{Ca} (n, \gamma) ^{49}\text{Ca}]$	8.72 m	3084.4
Na	100%	$[^{23}\text{Na} (n, \gamma) ^{24}\text{Na}]$	15.02 h	1368.6
K	17.9%	$[^{41}\text{K} (n, \gamma) ^{42}\text{K}]$	12.36 h	1524.7
Mn	99%	$[^{55}\text{Mn} (n, \gamma) ^{56}\text{Mn}]$	2.582 h	846.6

3. Results and Discussions

The measured concentrations of Al, Ca, Mg, V, Ti, Na, Mn, K, As, Br, La, Sm, Sc, U and Yb in the soil samples with the absolute standard deviations and the World median values of the elements estimated by Bowen [23] are presented in Table 4. In order to evaluate the accuracy of this method, three IAEA certified reference materials, (CRM) IAEA-soil-7, NBS-1633b Coal fly ash and IAEA-SL-1 were analyzed with the INAA-technique. The measured concentrations of different elements in the IAEA certified reference materials (NBS-1633b Coal fly ash, soil-7 and SL-1) are consistent with the certified values, with the exception of Na, Lu and Yb in IAEA-CRM-SL-1 which significantly deviated ($\pm 24.82\%$) from the certified values (Table 3). In Table 5, the mean concentrations, medians and the ranges of concentrations in the reactor surrounding soils were compared with that of the world soils. In almost all the samples, the lowest concentration is found except for Br, Yb and U which are above the corresponding range of the world. However, the maximum values of all the elements except Al, Na and Yb are found to be below the given world range. The mean value of Sc, K, U and Sm are higher than that of the world. In most cases the median values of the present experiment show close

agreement with that of the world median [23]. On the other hand, median values of measured elements except K, U, Sm, Sc and Na are below to the World median [23]. The concentrations of major elements (major nutrients) such as Na and K are in the range of 0.281-2.101%. On the regional scale, these soils are relatively rich in uranium and several other heavy metals such as potassium, vanadium and ytterbium.

According to Mattigot and Page, 1983 [27], concentrations of heavy metals are relatively higher in the nearby areas of reactor plants due to parent materials or secondary dispersion constitutes which can naturally cause the metal pollution. According to GLC (Great London Council) guidelines in Table 6, the measured concentration values are grouped into the following categories: uncontaminated, slightly contaminated, contaminated, heavily contaminated and unusually heavily contaminated. The pollution analysis of the soils of the present study revealed that Category II are slightly contaminated by manganese, Category I are not contaminated by arsenic and Category I and II are uncontaminated and slightly contaminated by vanadium (V). The highest concentrations of metals shown in Table 4 were observed in the samples of RFS 1, 7, 9, 10, 11, 13 and 14 (see Table1 for soil samples position).

The highest values of arsenic are observed in the samples RFS 6, 7 and 10, collected from areas at the north side of cooling tower, far from the decay tank and the west side of the underground tank, respectively. The lowest and highest concentrations of As are 1.58 and 7.13 mg kg⁻¹ respectively. The observed concentrations of As lie within the limiting values of world natural soil value (Mean 5 mg kg⁻¹) which is in agreement with the report by [28]. Furthermore, the concentration of As observed in the study area is comparable with the soil of coal-fired power plant (7.85±0.25 mg kg⁻¹) of South Carolina [29], different sites of Florida (0.02-7.01 mg kg⁻¹) [30], pre and post flooded area of New Orleans (3.69-23.36 mg kg⁻¹) [31] of the United States, King's Bridge Park sediments (5.07 mg/kg) of the Niagara river watershed at the U.S.- Canada border site [32]. However, the arsenic concentration is much lower than the value observed in the arsenic affected Singair area (44-60 mg kg⁻¹) in Bangladesh [33], Grand Duchy of Luxembourg (9-278 mg kg⁻¹) [34], Atlas steel mills (134 mg kg⁻¹ in oiled sample, 117 mg kg⁻¹ in de-oiled sample) of the city of welland, Ontario, Canada [32].

Vanadium is toxic to humans and other animals [35]. The greater part of the region does not seem to be polluted with Vanadium, except for the samples RFS 10, 12 and 14 collected from the west of the underground tank, the west side of the cooling tower pipe and the west side of reactor building, where V concentrations are higher than that of other areas. The concentration of V varies between 49.745-120.680 mg kg⁻¹, which is very much lower than the value observed (565 mg kg⁻¹) in the study conducted by [34]. However, the concentration of V is consistent with the Pre-Katrina and post-Katrina V concentrations (0.7-82.47 mg kg⁻¹) in soil in the area between the 17th Street Canal and Industrial Canal of Orleans Parish of United States [36, 37].

It is well known that Manganese is a very important metallic redox catalyst that governs the behavior of most trace metals in soils [38]. The concentrations of Mn in the samples RFS 1, 7 and 8 collected from the vicinity of the decay tank, far from the decay tank and the water treatment plant are higher than that of the other samples. The lowest and highest concentrations of Mn in the present study are 196.06 and 921 mg kg⁻¹, respectively whereas the range is 0.36-339.36 mg/kg in the soil of suburban

Table 3: Comparison of the measured elemental concentrations (in mg/kg) with the certified values

Certified Reference Materials	Element	This study	Certified values	RSD (%)
IAEA-SL-1	Ti	4642.3±733.4	5170±430	+10.2078
	V	143.23±9.309	170±15	+15.7475
	Na	1291.45±55.6	1700±100	+24.030
	K	12007±2425.4	14500±2100	+17.192
	Mn	3307±62.833	3460±160	+4.4218
	As	30.587±3.1	27.6±2.9	+10.82
	La	51.141±3.5	52.6±3.1	-2.77
	Lu	0.673±0.11	0.54±0.13	+24.63
	Sm	8.845±0.48	9.25±0.51	-4.37
IAEA-soil-7	Yb	4.269±0.43	3.42±0.65	+24.82
	Na	1921.7±42.32	2400±100	+19.921
	K	11116±800.35	12100±800	+8.1287
	Mn	639.42±44.75	631±27	-1.33367

areas of Varanasi, India [39], 26.9-5883 mg kg⁻¹ in Medvednica Nature Park, Croatia [40], 1.74-236 mg kg⁻¹ in different sites of Florida [30], 7.39-36.9 mg kg⁻¹ in the Al-Taif district of Saudi Arabia [41], 278.50-649.80 mg kg⁻¹ in Hangzhou, China [42]. In present study, the manganese concentration is much higher than the arsenic affected Singair area (22-80 mg kg⁻¹) in Bangladesh [33].

The observed range of samarium (Sm) in the soils of the area is 2.807-7.667 mg kg⁻¹, which is consistent with the values of world soils (2.3-7.7 mg kg⁻¹) [23, 43], sediments of King's Bridge Park (4.03 mg kg⁻¹) of the Niagara river watershed at the U.S.-Canada border site. However, the concentration of samarium (Sm) is greater than the value observed in the Atlas steel mills (0.48 mg kg⁻¹ in oiled sample, 0.58 mg kg⁻¹ in de-oiled sample) of the city of Welland, Canada [32]. Although the median value of Sm (6.342 mg kg⁻¹) exceeds the world median (4.5 mg kg⁻¹) [23], the areas are not contaminated by Sm. The highest values are found in the samples RFS 7 and 6 collected from the vicinity of the decay tank and the north side of the cooling tower, respectively.

Table 4: Concentrations of various elements in soils (in mg/kg, unless otherwise stated)

	Ti	Mg	V	Al	Ca
World median ^a	5000		90		15000
RFS-1	3480.9±250.624	1967.2±106.228	63.786±1.9773	47463±237.315	12133±1807.81
RFS-2	3049.5±195.168	1482.8±105.278	50.81±1.67673	39517±158.068	ND
RFS-3	2820.7±172.062	1554.6±79.2846	54.527±1.6358	39003±195.015	9072.7±1705.6
RFS-4	3440.9±230.200	2002.85±94.791	69.493±2.0425	51266±288.094	11430.35±1480.1
RFS-5	4061.1±288.338	2451.1±110.299	84.459±2.4493	63529±381.174	13788±1254.70
RFS-6	3772.7±279.179	1962.8±121.693	61.722±2.2837	46866±187.464	ND
RFS-7	2717.8±236.448	1718.7±96.2472	49.745±1.1441	38750±193.75	ND
RFS-8	3910±281.224	2070±116.1255	71.057±2.0425	52060±381.174	8680±1707.81
RFS-9	5260±320.052	3040±201.2583	89.808±1.9773	75330±288.094	10710±1827.31
RFS-10	5100±334.047	2970±111.0132	120.680±2.4493	78890±195.015	8190±807.21
RFS-11	5640±343.208	2980±107.0245	101.159±1.6767	75530±193.75	9350±1209.82
RFS-12	5340±298.034	3220±98.01254	104.827±2.2837	81090±237.315	10860±1877.51
RFS-13	6790±361.029	2740±119.1254	86.681±1.9773	75060±195.015	12510±1254.70
RFS-14	6700±362.025	3310±145.0125	105.319±2.0425	93930±158.068	13900±1802.81

^a Values from Ref.[23], N.D. Not Detected**Table 4:** Concentrations of various elements in soils (in mg/kg) (*Continued*)

	Na	K	Mn	As	Br	La
World median ^a	5000	14000	1000	6	10	40
RFS-1	ND	2805.7±530.277	920.99±4.60495	1.5850±0.4	2.713±0.89	16.533±5.21
RFS-2	ND	10058±382.204	328.81±4.27453	ND	ND	ND
RFS-3	5213.875±65.131	12871±592.066	503.15±5.0315	3.3440±0.7	5.826±1.25	30.370±12.22
RFS-4	3772.325±68.740	17763±515.127	507.88±15.7442	5.0550±1.1	5.749±1.63	35.981±5.212
RFS-5	6428.5±87.61265	19933±657.789	609.06±21.3171	4.1050±0.2	1.707±0.544	35.657±11.21
RFS-6	5754.15±75.849	18093±470.418	612.42±18.9850	6.0640±2.2	2.114±0.99	43.543±2.29
RFS-7	5792.175±92.196	18407±552.21	800.55±24.0165	6.2680±1.9	3.228±0.08	50.801±10.29
RFS-8	6054.7±125.5036	19184±633.072	777.03±25.6419	4.1640±1.2	3.138±1.23	42.316±12.31
RFS-9	7669.625±76.922	21012±693.396	546.18±18.0239	4.6410±1.6	4.886±1.10	36.025±6.21
RFS-10	5287.675±84.630	19737±532.899	ND	7.1330±2.1	3.143±1.47	46.374±11.91
RFS-11	6712.65±155.077	18398±754.318	646.23±32.3115	5.0870±1.1	5.406±1.23	35.376±7.25
RFS-12	6581.725±88.395	20372±855.624	672.46±38.3302	4.3840±1.6	2.813±0.54	37.849±10.21
RFS-13	3988.275±74.267	16396±705.028	698.22±32.1181	4.1920±0.9	7.015±2.32	43.757±10.33
RFS-14	4632.65±41.693	16093±675.906	196.06±12.3517	ND	ND	ND

^a Values from Ref.[23], N.D. Not Detected

Table 4: Concentrations of various elements in soils (in mg/kg) (*Continued*)

	Sm	Sc	Yb	U
World median ^a	4.5	7		2
RFS-1	2.807±0.245	8.234±2.782	2.328±0.4	1.9±0.3428
RFS-2	ND	9.169±3.214	0.159±0.06	0.016±0.0098
RFS-3	5.016±1.042	6.805±2.451	2.721±0.7	0.953±0.246
RFS-4	5.931±2.275	10.140±0.789	2.667±1.1	2.743±0.486
RFS-5	5.637±1.245	10.545±1.245	5.178±0.9	5.371±2.1031
RFS-6	7.022±1.127	12.501±0.847	2.007±0.6	2.331±0.5428
RFS-7	7.667±2.242	12.795±3.542	4.036±0.2	4.705±1.785
RFS-8	6.730±1.225	11.481±2.461	5.271±1.2	6.170±2.432
RFS-9	6.003±2.149	12.977±1.125	2.714±0.5	3.189±0.285
RFS-10	6.884±1.845	12.561±0.987	1.987±0.8	2.345±0.5452
RFS-11	6.148±1.245	10.034±1.52	5.365±0.56	6.356±2.036
RFS-12	6.536±2.234	11.354±2.14	4.403±1.2	5.237±1.932
RFS-13	6.935±2.115	8.893±3.21	2.606±1.1	3.113±0.985
RFS-14	ND	9.925±3.25	0.112±0.03	0.013±0.009

^a Values from Ref.[23], N.D. Not Detected

The mean and median values (10.53 and 10.34 mg kg⁻¹) of scandium in the study is exceed over the mean (7 mg kg⁻¹) and median (7 mg kg⁻¹) value of the world soil, the soil of King's Bridge Park (5.5 mg kg⁻¹) and in the sediments of Atlas steel mills (0.31 mg kg⁻¹ in oiled sample, 0.54 mg kg⁻¹ in de-oiled sample) of the city of welland, Canada [32], although the observed concentrations of Sc (6.805 to 12.977 mg kg⁻¹) lie within the limiting value of world natural soil (0.5-45 mg kg⁻¹) [23,43]. The highest value of Sc was found in samples RFS 6, 9, and 10 collected from the vicinity of the cooling tower, the southwest boundary wall of reactor and the underground tank, respectively. Severe pollution is observed for ytterbium (Yb) in the sample RFS 5, 8 and 11 collected from the vicinity of the over- head tank, water treatment plant and garden beside the lemon tree. The greater part of the region is not however appreciably polluted. The highest value of Yb observed in samples from the garden beside the lemon tree is nearly double the world mean value for natural soil [23]. The lowest and highest concentrations of Yb are 0.112 and 5.365 mg kg⁻¹ respectively. The range of Yb for this region is above the range of world soil (1.8 - 3.4 mg kg⁻¹) and the mean value (2.96 mg kg⁻¹) of the study area exceeded the value found in the soil of King's bridge park (1.87 mg kg⁻¹) and Atlas steel mills (<1 mg kg⁻¹) of Canada [32].

The concentrations of Mg, Ca and Br are relatively low in the study area. The obtained means and medians for these elements are below the means and medians of world soils. The maximum values of Mg are found in samples RFS 9, 12, 14 collected from the south west of reactor boundary, west side of the cooling tower pipe and the west side of reactor building respectively. The higher values are found in samples RFS 5, 10, 14 for Ca and samples RFS 3, 11, 13 for Br where the locations are mentioned in Table1. The lowest and highest concentration varies between 0.1483 and 0.331% for Mg, 0.82 and 1.38% for Ca, whereas this limit observed very lower in the soil of Al Taif district of Saudi Arabia, 0.0115-0.0118% for Mg and 0.0547-0.0889% for Ca [41]. The lowest and highest concentrations for

Br are 1.707 and 5.826 mg kg⁻¹. The measured Br of this area (3.97 mg kg⁻¹) is considerably lower than the soil of the King's bridge park (10.37 mg kg⁻¹) of Canada [32]. The soil of this area is not contaminated by Mg, Ca and Br (Table 4,5,6).

Table 5: Range, mean and median value for reactor surrounding soils and world soils. All values are in (mg/kg) unless otherwise specified

Element	Soil Surround the TRIGA MARK II Research Reactor and nearby nuclear installations			World		
	Range	Mean	Median	Range	Mean ^c	Median ^a
Al (%)	3.87-9.39	6.13	5.77	1-6.4 ^d	7.13	NA
As	1.58-7.13	4.80	4.84	0.1-40 ^e	5	6
Br	1.70-5.82	3.97	3.18	5-40 ^b	5	10
Ca (%)	0.81-1.37	0.97	1.07	NA-NA	1.37	1.5
La	16.53-50.80	37.88	36.93	13.9 - 56.3 ^b	40	40
Mg (%)	0.14-0.33	0.23	0.22	NA-NA	0.63	NA
Mn	196.06-920.99	601.46	612.42	100-4000 ^c	850	1000
Na (%)	0.37-0.76	0.56	0.57	0.07-0.74 ^d	0.63	0.50
Sc	6.80-12.97	10.53	10.34	0.5-45 ^b	7	7
Sm	2.80-7.66	6.10	6.34	2.3-7.7 ^b	4.5 ^a	4.5
Ti (%)	0.27-0.67	0.44	0.39	0.1-0.9 ^b	0.46	0.50
U	0.01-6.35	3.17	2.98	0.79- 11 ^b	1	2
V	49.74-120.68	79.57	77.75	3-500 ^a	100	90
Yb	0.11-5.36	2.96	2.69	1.8 - 3.4 ^b	3 ^a	NA
K(%)	0.28-2.10	1.65	1.82	NA	1.40 ^b	1.40

^aRef.[23], ^bRef.[43], ^cRef.[1], ^d Ref.[25], ^e Ref.[28], NA : No available data.

The lowest and highest concentrations of Ti are 0.272% and 0.68 % which is within the limiting values of world soil (0.1-0.9%). Similar trend is observed for La in the present study. The concentration range for La in the present study is 16.53-50.80 mg kg⁻¹ which is also comparable with world soil (13.9-56.3 mg kg⁻¹) in Table 5. In addition the amount of La observed in the King's bridge park soil (15.67 mg kg⁻¹) [32] is comparable to the lowest limit of the present study. The highest values of Ti are observed in samples of RFS 13 and 14 from the north-east corner of the reactor building and the west side of the reactor building, respectively. The highest values of La are found in samples RFS 7 and 10 collected from the vicinity of the decay tank and the underground tank. The region seems to be not contaminated by Ti and La (Table 4).

Table 6: The classification of soils based on G.L.C guidelines (elements values in mg/kg unless otherwise specified)

Element	Range in reactor surrounding soils	Mean in reactor surrounding soils	Average soil composition ^a	G.L.C. guidelines ^b				
				I	II	III	IV	V ^c
Mn	196.06-921.00	601.50	320	500	1000	2000	1%	>1%
As	1.59-7.13	4.80	7.5	30	50	100	500	>500
V	49.75-120.70	79.57	57	200	200	500	2500	>2500

^a Average soil composition from Ref.[26], ^bGLC (Great London Council), ^c Ref.[46], I=Typical values for uncontaminated soils, II=Slight contamination, III=Contaminated, IV=Heavy contamination, V =Unusually heavy contamination

Heavy pollution by K is observed in the soil surrounding the reactor, where its median value (1.8%) exceeds the median (1.4%) of world soil [23]. All samples except the sample of RFS 1, 2 and 3 have a higher value of K (Table 4). The observed limiting value for K of this area 0.28-2.1% is remarkably greater than the corresponding values measured in Saudi Arabia 0.03-0.04% [41]. The vicinity of the over head tank, south west of reactor boundary and the west side of the cooling tower pipe are polluted by K. This area is totally not contaminated by Na as it's limiting values (0.37-0.76%) are comparable with the range of world soils (0.07-0.74%) [23] and the soil of Saudi Arabia (0.01-0.24%) [41]. The few sites of the study area are slightly contaminated by Al. The concentrations obtained for Al (3.8-9.3%) are comparable with the world soil (1.0-6.4%) with the exception of RFS 12 (8.1%) and RFS 14 (9.3%). In addition, a significant difference is shown for the value of Al (0.01-1.2%) in Finland [44].

The maximum value found for Uranium is nearly three times higher than the value for natural soil estimated by [23]. The lowest and highest concentration varies from 0.01 to 6.35 mg kg⁻¹ (Table 5), which is remarkably higher than the background level observed in Belgium (1 mg kg⁻¹) [45], King's bridge park area (1.3 mg kg⁻¹) and the Atlas Steels (0.65-0.81 mg kg⁻¹) of Canada [32]. In addition the sediments of the Quirke (24.3±12.2 mg kg⁻¹), McCarthy (44.6±9.9 mg kg⁻¹), McCabe (38.6 mg kg⁻¹), and Elliot (133±46.1 mg kg⁻¹) lakes of the Ontario in Canada showed significant difference with the uranium observed in the study area as well as the control lake Semiwite (4.1 mg kg⁻¹) [24]. The higher values of Uranium are observed in the samples of RFS 5, 8, 11 collected from the vicinity of the overhead tank, water treatment plant and the vegetable garden. The chief sources are primarily nuclear activities of the coolant tower, underground tank and overhead tank beside the vegetable garden. It seems that the larger parts of the region are significantly polluted by Uranium.

4. Conclusion

The study clearly showed that in certain location, the soil samples have higher values of concentration(mean, median and range) compared to that of the world which indicating that the soils of these sites have been contaminated by the nuclear activities of the reactor and other nearby installations. The measured results show that surface soils in this region are polluted to a certain extent with several metals, for example K, U, Sm, Sc, Al, Na, and Yb. The heavy metal (U and Yb) pollution is particularly pronounced in areas near the overhead tank, water treatment plant and the vegetable garden, K everywhere. The contamination level is relatively high for Mn in decay tank and water treatment plant area and V in the vicinity of the underground tank, the cooling tower and the west side

of reactor building area. However, there are no available scientific studies that indicate that there are harmful effects of this pollution on the human population or other biological species. Still, large amounts of vegetables and fruits are harvested in the gardens and farmlands beside the TRIGA reactor and are regularly taken by the people of the surrounding area. It is therefore, necessary to determine the levels of heavy metals and other trace elements in soils, and uptake by vegetations and other biological species and take appropriate steps to protect the people from the health hazard of the excessive levels of metals. The degree of soil pollution around the TRIGA reactor poses a risk to human health, because it may pollute the fresh water sources that supply drinking water to the institute and the nearby areas. Further studies should be carried out to understand the full-phase level of contamination in the water and vegetation of the nearby areas of the reactor plant.

5. Acknowledgement

The authors would like to thank Dr. Michelle Shaw, Research Consultant, ACEnet, Memorial University, for her assistance to improve the manuscript. We also gratefully acknowledge Dr. Feroza Akter and reactor operational staff of Bangladesh Atomic Energy Research Establishment (AERE), Savar, Dhaka, for their kind cooperation during the experiment

6. References

1. Coskun, M., E. Steinnes, M. Viladimirovna, Frontasyeva, T. E. Sjobakk and S. Demkina, 2006. Heavy metal pollution of surface soil in the Thrace region, Turkey. *Environ. Monit. and Assess.*, 119: 545–556.
2. Kabata-Pendias, A., 2004. Soil–plant transfer of trace elements—an environmental issue. *Geoderma*, 122: 143-149.
3. Djingova, R. and I. Kuleff, 2000. Instrumental Techniques for Trace Analysis. In: *Trace elements-Their Distribution and Effects in the Environment*, Markert, B. and K. Friese, (Eds.). Elsevier, Amsterdam, pp:137–185.
4. Bermea, O. M., A. H. Alvarez, I. Gaso and N. Segovia, 2002. Heavy metal concentrations in surface soil from Mexico City. *Bull. Environ. Contam. Toxicol.*, 68: 383– 388.
5. Kalantari, M. R., M. Shokrzadeh, A. G. Ebadi, C. Mohammadzadeh, M. I. Choudhary and A. Rahman, 2006. Soil pollution by heavy metals and remediation (Mazandaran-Iran). *J. Applied Sci.*, 6: 2110-2116.
6. McLaughlin, M.J., 2001. Bioavailability of Metals to Terrestrial Plants. In: *Bioavailability of Metals in Terrestrial Ecosystems. Importance of Partitioning for Bioavailability to Invertebrates, Microbes and Plants*, Allen, H.E. (Eds.), SETAC Press, Pensacola, FL, pp: 39-68.
7. Uchida, S., K. Tagami and I. Hirai., 2007. Soil-to-plant transfer factors of stable elements and naturally occurring radionuclides: (2) Rice collected in Japan. *J. Nucl. Sci. Technol.*, 44: 779–790.

8. Rana, M.S., M.A. Halim, S. Safiullah, M.M. Mollah and M.S. Azam *et al.*, 2009. Removal of heavy metal from contaminated water by biopolymer crab shell chitosan. *J. Applied Sci.*, 9: 2762-2769.
9. Naidu, G. R. K., N. Trautmann, S. Zaunar, T. Balaji and K.S. Rao, 2003. Multielemental analysis of soils by instrumental neutron activation analysis. *J. Radioanal. Nucl. Chem.*, 258:421-425.
10. Steinnes, E., R.O. Allen, H.M. Petersen, J.P. Rambek and P. Varskog, 1997. Evidence of large scale heavy-metal contamination of natural surface soil in Norway from long-range atmospheric transport. *Sci. Total. Envir.*, 205:255–266.
11. Steinnes, E., N. Lukina, V. Nikonov, D. Aamlid and O. Royset, 2000. A gradient study of 34 elements in the vicinity of a copper-nickel smelter in the Kola Peninsula. *Environ. Monit. Assess.*, 60: 71–88.
12. Kabata-Pendias, A. and H. Pendias., 2000. Trace Elements in Soils and Plants. 3rd Edn., CRC Press Inc., Boca Raton, USA.
13. Donisa, C., R. Mocanu, E. Steinnes, and A. Vasu, 2000. Heavy metal pollution by atmospheric transport in natural soils from the northern part of eastern Carpathians. *Water, Air, and Soil Pollution*, 120: 347–358.
14. Fytianos, K., G. Katsianis, P. Triantafyllou and G. Zachariadis, 2001. Accumulation of heavy metals in vegetables grown in an industrial area in relation to soil. *Bull. Environ. Contam. Toxicol.*, 67: 423–430.
15. Sanghi, R. and K. S. Sasi, 2001. Pesticides and heavy metals in agricultural soil of Kanpur, India. *Bull. Environ. Contam. Toxicol.*, 67: 446–454.
16. Lin, Y. P., T. P. Teng and T. K. Chang, 2002. Multivariate analysis of soil heavy metal pollution and landscape pattern in Changhua country in Taiwan. *Landscape and Urban Planning*, 62:19–35.
17. Einax, J.W. and U. Soldt, 1998. Multivariate geostatistical analysis of soil contaminations. *Fresenius J. Anal. Chem.*, 361: 10–14.
18. Plant, J., D. Smith, B. Smith and L. Williams, 2001. Environmental geochemistry at the global scale. *Appl. Geochem.*, 16: 1291–1308.
19. Facchinelli, A., E. Sacchi, and L. Mallen, 2001. Multivariate statistical and GIS-based approach to identify heavy metal sources in soils. *Environ. Pollut.*, 114: 313–324.
20. Brus, D. J., J. J. de Gruijter, D. J. Walvoort, F. de Vries, J. J. Bronswijk, P. F. Romkens and W. de Vries, 2002. Mapping the probability of exceeding critical thresholds for Cadmium concentrations in soils in the Netherlands. *J. Environ. Qual.*, 31: 1875–1884.
21. Aslibekian, O. and R. Moles, 2003. Environmental risk assessment of metals contaminated soils at silvermines abandoned mine site, Co. Tipperary, Ireland. *Environ. Geochem. Health*, 25: 247-266.
22. Mariotti, C., 1998, Legislative aspect and risk assessment for contaminated sites in Italy: the UNICHIM approach. In: Risk Assessment and Risk Management of Contaminated Sites in

-
- Europe, Guercia, F. and C. Mariotti, (Eds.). Proceeding of the International Workshop, ANPA, Rome, pp. 69–80.
23. Bowen, H. J. M., 1979. *The Environmental Chemistry of the Elements*. Academic Press. London.
24. Clulow, F.V., N. K. Dave, T. P. Lim and R. Avadhanul, 1998. Radionuclides (lead-210, polonium-210, thorium-230, and-232) and thorium and uranium in water, sediments, and fish from lakes near the city of Elliot Lake, Ontario, Canada. *Environ. Pollut.*, 99: 199-213.
25. Bear, F. E., 1964. *Chemistry of the Soil*. 2nd Edn., Reinhold, New York.
26. Connor, J. J. and H. T. Shacklette with section on field studies by R. J. Ebans, J. A. Erdman, A. T. Miesel, R. R. Tidball, and H. A. Tourtelot, 1975. Back-ground geochemistry of some soils, plants and vegetables in the conterminous United States. In: *United States Geological Survey and Research Professional Papers*, 574-F, p.164.
27. Mattigot, S.V. and A. L. Page, 1983. Assessment of Metal Pollution in Soils. In: *Applied environmental geochemistry*, Thorton (Eds.). Academic Press, London, pp: 355–395.
28. Alloway, B. J., 1994. *Heavy Metals in Soils*. 2nd Edn., Springer, New York.
29. Peltier, G. L., M. S. Wright, W. A. Hopkins, J. L. Meyer, 2009. Accumulation of trace elements and growth responses in *Corbicula fluminea* downstream of a coal-fired power plant. *Ecotoxicol. Environ. Safety*, 72:1384–1391.
30. Chen, M., L. Q. Ma and W.G. Harris, 1999. Baseline concentrations of 15 trace elements in Florida surface soils. *J. Environ. Qual.*, 28:1173-1181.
31. Rotkin-Ellman, M., G. Solomon, C. R. Gonzales, 2010. Arsenic contamination in New Orleans soil: Temporal changes associated with flooding. *Environ. Res.*, 110: 19–25.
32. Dickman, M., G. Rygiel, 1996. Chironomid Larval deformity frequencies, mortality and diversity in heavy metal contaminated sediments of a Canadian riverine wetland. *Environ. Int.*, 22: 693-703.
33. Islam, M.S., M.A. Halim, S. Safiullah, M.S. Islam, M.M. Islam, 2009. Analysis of organic matter, iron and manganese in soil of arsenic affected Singair area, Bangladesh. *Res. J. Environ. Toxicol.*, 3: 31-35.
34. Horckmans, L., R. Swennena, J. Deckers and R. Maquil, 2005. Local background concentrations of trace elements in soils: A case study in the grand duchy of Luxembourg. *Catena*, 59: 279–304.
35. Perez-Benito, J. F., 2006. Effects of chromium (VI) and vanadium (V) on the lifespan of fish. *J. Trace Elem. Med. Biol.*, 20: 161–170.
36. Wang, G., H.W. Mielke, V. Quach, C. Gonzales and Q. Zhang, 2004. Determination of polycyclic aromatic hydrocarbons and trace metals in New Orleans soils and sediments. *Soil Sediment Contamination*, 13: 313–327.

37. Su, T., S. Shu, H. Shi, J. Wang, C. Adams and C. Witt, 2008. Distribution of toxic trace elements in soil/sediment in post-katrina new orleans and the louisiana Delta. *Environ. Pollut.*, 156: 944–950.
38. Bartlett, R. J. , 1998. Characterizing Soil Redox Behavior. In: *Soil Physical Chemistry*, Sparks, D. L. (Eds.). CRC Press, Boca Raton, FL, pp: 371– 391.
39. Sharma, R.K., M. Agrawal and F. Marshall, 2007. Heavy metal contamination of soil and vegetables in suburban area of Varanasi, India. *Ecotoxicol. Environ. Safety*, 66: 258–266.
40. Jelaska, L. S., M. Blanusa, P. Durbesic and S.D. Jelaska, 2007. Heavy metal concentrations in ground beetles, leaf litter, and soil of a forest ecosystem. *Ecotoxicol. Environ. Safety*, 66: 74–81.
41. Mohamed, A. E, M. N. Rashed and A. Mofty, 2003. Assessment of essential and toxic elements in some kinds of vegetables. *Ecotoxicol. Environ. Safety*, 55: 251–260.
42. Chen, T., X. Liu, M. Zhu, K. Zhao, J. Wu, J. Xu and P. Huang, 2008. Identification of trace element sources and associated risk assessment in vegetable soils of the urban-rural transitional area of Hangzhou, China. *Environ. Pollut.*, 151: 67-78.
43. Kabata-Pendias, A. and H. Pendias, 1984. *Trace Element in Soils and Plants*. CRC Press, New York.
44. Lukkari, T., M. Taavitsainen, A. Vaisanen, J. Haimi, 2004. Effects of heavy metals on earthworms along contamination gradients in organic rich soils. *Ecotoxicol. Environ. Safety*, 59: 340–348.
45. Ruffykiri, G., L. Huysmans, J. Wannijn, M. Van Hees, C. Leyval and I. Jakobsen, 2004. Arbuscular mycorrhizal fungi can decrease the uptake of uranium by subterranean clover grown at high levels of uranium in soil. *Environ. Pollut.*, 130: 427-436.
46. Skordas, K. and A. Kelepertsis, 2005. Soil contamination by toxic metals in the cultivated region of Agia, Thessaly, Greece. Identification of sources of contamination, *Environ. Geol.*, 48: 615-624.